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Associations between multiple ambient air pollutants, genetic risk, and incident mental disorders: An interaction study in the UK population

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HIGHLIGHTS

G R A P H I C A L A B S T R A C T

- Air pollutants were associated with a higher risk of incident mental disorders.
- Risk of incident schizophrenia, depression, and anxiety were assessed specifically.
- Genetic risks of schizophrenia, depression and anxiety did not alter the associations.
- Air pollution control is important for the prevention of mental disorders.

Associations between multiple ambient air pollutants, genetic risk, and incident mental disorders: an interaction study in the UK population

Air Pollution Score Associations between air pollution and risk of incident mental disorders (N = 252,376)

weighted	Mental disor	ders	Schizophre	nia	Depressio	on	Anxiety	
Per IQR increase	Hazard Ratio	P value						
2.5 absorbance PM2.5 absorbance	1.02 (1.01, 1.04)	0.002	1.18 (1.05, 1.33)	0.006	1.01 (0.98, 1.04)	0.483	1.02 (1.00, 1.05)	0.072
M2.5-10 PM2.5	1.06 (1.04, 1.08)	< 0.001	1.14 (0.98, 1.34)	0.085	1.04 (1.01, 1.08)	0.009	1.07 (1.03, 1.10)	< 0.001
PM2.510	1.00 (0.99, 1.01)	0.968	1.01 (0.92, 1.10)	0.904	1.00 (0.99, 1.02)	0.588	1.00 (0.99, 1.02)	0.657
PM ₁₀	1.00 (0.99, 1.01)	0.613	1.02 (0.91, 1.13)	0.751	1.01 (0.99, 1.03)	0.370	1.02 (1.00, 1.04)	0.131
NO ₂	1.06 (1.04, 1.08)	< 0.001	1.29 (1.09, 1.52)	0.003	1.04 (1.00, 1.08)	0.030	1.07 (1.03, 1.11)	< 0.001
NOx	1.04 (1.03, 1.06)	< 0.001	1.16 (1.04, 1.29)	0.006	1.03 (1.00, 1.05)	0.047	1.04 (1.02, 1.07)	0.001
Air pollutants score	1.06 (1.04, 1.07)	< 0.001	1.24 (1.08, 1.43)	0.003	1.04 (1.01, 1.07)	0.022	1.06 (1.03, 1.10)	< 0.001



Air pollutants were associated with an increased risk of mental disorders, including schizophrenia, depression and anxiety. There is no interaction between air pollution and the genetic risk of schizophrenia, depression or anxiety in corresponding incident disorders. Results show important implications for air pollution control to prevent mental disorders.

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ABSTRACT

Mental disorders can be triggered by genetic and environmental risk factors. Limited studies have explored the effects of long-term exposure to air pollution on mental disorders, and most of the studies have focused on individual air pollutants. This study aimed to examine the relationship between long-term exposure to multiple air pollutants and incident mental disorders, including depression, anxiety, and schizophrenia, and whether the associations were affected by genetic susceptibility. Participants in the UK Biobank with no history of mental disorders were followed from baseline (2006 to 2010) to October 31st, 2022. Cox regression was applied to evaluate the correlation between $PM_{2.5}$ absorbance, $PM_{2.5}$, $PM_{2.5-10}$, PM_{10} , NO_2 , and NOx and any or specific mental disorders. Additive and multiplicative scales were used to measure the interaction between air pollution and schizophrenia polygenic risk score (PRS), depression PRS, or anxiety PRS on specific mental diseases. After a median of 13.36 years of follow-up on 252,376 participants, we observed per interquartile increase of $PM_{2.5}$ absorbance (0.32 per meter), $PM_{2.5}$ (1.28 µg/m³), NO₂ (10.08 µg/m³), and NO_x (16.78 µg/m³) were related to a

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2–6 % higher risk of incident mental disorders. The HR (95 % CI) of incident mental disorder for the 2nd, 3rd, and 4th quartile of the air pollution score were 1.05 (1.01–1.18), 1.13 (1.09–1.18), and 1.14 (1.09–1.19), respectively, in comparison to the lowest level of the score. Per interquartile increase in the air pollution score was associated with a 6 %, 24 %, 4 %, and 6 % higher risk of incident mental disorders, schizophrenia, depression, and anxiety, respectively. No interaction between air pollution and genetic risk of schizophrenia, depression or anxiety on corresponding incident disorders was observed. These findings emphasize the importance of implementing air pollution control standards to decrease the burden of mental disorders.

1. Introduction

Systematic reviews have demonstrated that mental disorders introduce an increased risk of premature mortality (Plana-Ripoll et al., 2019). In terms of disability-adjusted life years (DALYs), the global number of DALYs due to mental disorders increased by 55 % between 1990 and 2019, moving from the 13th to the 7th leading cause (GBD, 2019). Mental disorders contribute to a significant burden of disease worldwide (Adorjan and Falkai, 2019). In 2019, depressive disorders, anxiety disorders, and schizophrenia were recognised as the top three specific mental disorders (GBD, 2019). Identifying risk factors for mental disorders and intervening effectively would, therefore, be an important public health priority to reduce their risks and burden on healthcare and economics.

Ambient air pollution is one of the most significant health threats (World Health Organization, 2023). Excess risks of cardiovascular disease, respiratory disease and cancers caused by ambient air pollution have been widely reported (Landrigan et al., 2018; Cohen et al., 2017; Figueres et al., 2018). In recent years, air pollutants' effects on the central nervous system have been increasingly recognised (Buoli et al., 2018). Mechanisms indicated that different air pollutants could induce neuroinflammation and oxidative stress, a plausible pathogenesis of mental disorders, through various pathways (Hurley and Tizabi, 2013; Szymkowicz et al., 2023; Lyons et al., 2024; Lundberg and Weitzberg, 2022; Adar et al., 2014). Distinct mental disorders may involve different biological pathways and exhibit varying susceptibility to air pollutants (Thomson, 2019; Dantzer et al., 2008; Hasler, 2010; Marsman et al., 2013; McCutcheon et al., 2023; Braithwaite et al., 2019). Recent studies have investigated the relationship between air pollution and mental well-being. Nevertheless, there was a lack of achieved consensus on associations between air pollution and certain mental disorders, such as depression, anxiety or schizophrenia, due to variations in outcome identification and study design (Braithwaite et al., 2019; Fan et al., 2020). Ambient air pollution is a complex mixture of particulate matter and gaseous pollutants. However, previous studies have primarily discussed the individual effects of single air pollutants, especially particulate matter (Braithwaite et al., 2019; Zeng et al., 2019; Nobile et al., 2023). Exploring the compound effects of this mixture on incident mental disorders will be crucial to gaining a more comprehensive understanding of how air pollution affects mental well-being.

Recent studies have suggested the existence of an interplay between genetic predisposition and environmental factors in the risk of chronic disease (Peters et al., 2021). It is well known that the occurrence and development of mental disorders are influenced by genetic and environmental risk factors (Assary et al., 2018). Genome-wide association studies (GWAS) have identified hundreds of potential genetic risk variants of several mental disorders, such as schizophrenia, depression, and anxiety (Howard et al., 2019; O'Donovan et al., 2008; Li et al., 2024). These genetic determinants would preliminarily identify at-risk individuals in the pre-disease stage. Nevertheless, it is unclear to what extent genetic predisposition affects the relationship between concurrent exposure to certain air contaminants and the development of specific mental disorders. Consequently, there is a need to explore the combined effects of genetic and environmental risk factors to improve the risk assessment for mental disorders.

The primary objective of this study is to investigate the association

between long-term exposure to multiple air pollutants and any or specific mental disorders, including schizophrenia, depression and anxiety, with data from a national longitudinal cohort in the UK. Additionally, we aim to examine whether genetic predisposition interacts with air pollution and thus alters the associations between air pollution and specific mental disorders.

2. Methods

2.1. Study population and data source

The study's data was sourced from the UK Biobank, a nationwide prospective cohort of over 500,000 people in the UK who were initially recruited between 2006 and 2010 when they were aged 40 to 69. The UK Biobank collected comprehensive information about participants through questionnaires, physical assessments, and biological sampling, allowing for a thorough understanding of their phenotypic characteristics. Participants' genotypic information was collected, and longitudinal health outcomes were obtained through data linkage from various sources, including self-reports, primary care records, hospital admissions, cancer register, and death records.

We collected participants with information on air pollutants, excluded participants without complete information on covariates and excluded those who had any mental disorders at baseline. The flowchart for participant selection was shown in Fig. 1. We followed the eligible participants from the date of baseline assessment until the date of death, the earliest date of diagnosis of mental disorders, or 31st October 2022, whichever came first.

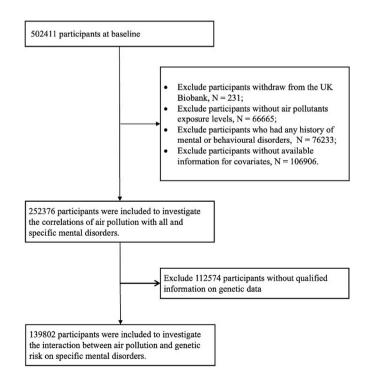


Fig. 1. Flowchart of study population selection.

2.2. Estimation of air pollutants and air pollution score

Estimates of ambient air pollutants for each residential baseline address with a 100 m*100 m resolution, including PM2.5 absorbance (a proxy for elemental carbon) (Eeftens et al., 2012a), PM_{2.5}, PM_{2.5-10}, PM₁₀, NO₂, and NOx were assessed by the UK Biobank, using a Land Use Regression (LUR) model developed as part of the European Study of Cohorts for Air Pollution Effects (ESCAPE, http://www.escapeproject. eu/) for the year 2010 (Eeftens et al., 2012b). The annual average concentrations of these six pollutants in 2010 were used to determine the air pollutant exposure levels for each individual. The LUR model was widely employed in high-impact research (Wolf et al., 2021; Cesaroni et al., 2014; Khomenko et al., 2021). We did not use air pollutant estimates provided by the UK Department for Environment, Food and Rural Affairs (DEFRA), because the LUR model was cost-effective and provided higher resolution data compared to the DEFRA's model (100 m*100 m vs 1 km*1 km) (Hoek et al., 2008; UK Air, 2022). In addition, the LUR model was able to generate exposure estimates for particulate matter components such as absorbance and ultrafine particulate matter, which were not available in DEFRA's data, and thus served the purpose of our research to better understand how these contaminants affect health (de Hoogh et al., 2013). Additionally, we constructed weighted air pollution scores to measure the combination effects of air pollutants. The score was calculated by adding the concentrations of each air pollutant and weighting them using the multivariable-adjusted risk estimates (β coefficients). These β coefficients originated from the adjusted associations between each air pollutant and any or specific mental disorders conducted in the primary analysis (detailed in Section 2.5). Previous research on UK Biobank data has made considerable use of this score to evaluate the link between concurrent exposure to cumulative air pollutants and the risk of diseases, e.g. chronic heart failure and rheumatoid arthritis (Zhang et al., 2023; Wang et al., 2021).

$$\begin{aligned} \text{Air pollution score} &= \left[\beta_{PM_{2.5} \text{ absorbance}} \times PM_{2.5 \text{ absorbance}} + \beta_{PM_{2.5}} \times PM_{2.5} \right. \\ &+ \beta_{PM_{2.5-10}} \times PM_{2.5-10} + \beta_{PM_{10}} \times PM_{10} + \beta_{NO_2} \times NO_2 \\ &+ \beta_{NO_x} \times NO_x \right] \times \left[6 / (\beta_{PM_{2.5} \text{ absorbance}} + \beta_{PM_{2.5}} + \beta_{PM_{2.5-10}} \\ &+ \beta_{PM_{10}} + \beta_{NO_2} + \beta_{NO_x}) \right] \end{aligned}$$

2.3. Outcome and covariate

The study identified the first instance of the incident mental disorders using the International Classification of Disease-10 (ICD-10), F00-F99, determined by self-reports, primary care records, hospital admissions, and death records (Liu et al., 2024; Gao et al., 2017). At the baseline assessment (and at follow-up assessments if they attended them), participants were asked to self-report all doctor-diagnosed health conditions. Their responses were further checked and verified by a nurse through a verbal interview. All self-reported diagnosed conditions and diagnoses from other sources were coded to the 3-character ICD-10 codes where appropriate (https://biobank.ndph.ox.ac.uk/showcase/re fer.cgi?id=593). Given the high prevalence in the general population (GBD, 2019; Liu et al., 2024), we also examined three specific mental disorders, including schizophrenia (F20-F29) (Gao et al., 2017; Newbury et al., 2021), depression (F32) (Fu et al., 2022), and anxiety (F40-F41) (Liu et al., 2024), were also examined. The study used a directed acyclic graph (DAG, https://dagitty.net/dags.html#) to identify the minimal sufficient adjustments. Several potential covariates, including sex, ethnicity, age at the baseline, educational attainment, duration of residence at the current address, family income before taxes, employment status, Townsend deprivation index, level of physical activity, and percentage of green space coverage within a 300-m radius of the residence were considered (Supplementary Fig. 1). We used the Townsend deprivation index as the small area-level socioeconomic indicator. The deprivation index was calculated based on output areas from the most recent national census before participants were enrolled in the UK Biobank. These output areas have measures including households car ownership; households overcrowding; households owner occupation and persons unemployed. A higher deprivation score indicates greater socioeconomic deprivation (Foster et al., 2018; Blane et al., 1987).

2.4. Polygenetic risk score for schizophrenia, depression and anxiety

The UK Biobank offers a comprehensive repository of genetic data on research subjects. Genotyping was conducted using the Applied Biosystems UK BiLEVE Axiom Array by Affymetrix, and Applied Biosystems UK Biobank Axiom Array covered 95 % of the gene sites (Bycroft et al., 2018). For the remaining 5 % of unique single-nucleotide polymorphisms (SNPs), imputation was performed using SHAPEIT3 and IMPUTE2 algorithms. The Haplotype Reference Consortium data was used as the main reference panel for imputation (Bycroft et al., 2018). We determined the polygenic risk scores (PRS) of schizophrenia, anxiety, and depression via the summary statistics from meta-analyses of GWAS excluding UK Biobank participants (Howard et al., 2019; Wray et al., 2018). This is to avoid sample overlap between base (i.e., the summary statistics) and target (UK Biobank) datasets. Summary statistics of these meta-analyses are available in the Psychiatric Genomics Consortium (schizophrenia: scz2022; anxiety: anx2016; depression: mdd2018 noUKBB) (Trubetskoy et al., 2022; Otowa et al., 2016; Adams, 2022). We excluded SNP with a low minor allele frequency of <0.005and a low imputation accuracy of <0.1. We removed participants with non-European ancestry and those who are genetically correlated. We used all cases of specific mental disorders up until the end of follow-up to derive the PRS, with PRSice 2.0 program (Euesden et al., 2015). The clumping threshold was set as P = 1, linkage disequilibrium score $r^2 =$ 0.25, and distance threshold = 250 kb (Shen et al., 2020). P-value cutoffs for thresholding SNPs in the summary statistics were set at 0.0005, 0.001, 0.005, 0.01, 0.05, 0.1, 0.5, and 1. We used the PRS derived from the best-fit *p*-value cut-off threshold, which provided the highest R-squared value for prediction, in our genetic interaction analysis.

2.5. Statistical analysis

Continuous variables were reported as means with standard deviations (SD) or median with interquartile range (IQR). Categorical variables were reported as frequencies and percentages. We used the Cox regression model to evaluate the correlation of individual air pollutants and air pollution score with all or specific incident mental disorders, including schizophrenia, depression and anxiety, respectively. Based on the quartiles of the exposures, we sorted the individuals into four groups; the reference group was the group with the lowest exposure (the first quartile). Additionally, for each IQR increase in the exposures, the associated hazard ratio (HR) was computed. We conducted analyses using both crude HR and HR adjusted for potential covariates. The crude model was controlled for age (in a linear term) and sex. The adjusted model was controlled for ethnicity (White and non-White), annual family income before tax (< £18,000, £18,000 to £30,999, £31,000 to £51,999, £52,000 to £100,000, and £100,000 above), the highest educational qualification (University or college, A-level, GSCE and other), employment status (unemployed, employed and retired), duration of residence at the current address (in quintiles) and the Townsend deprivation index (in quintiles), physical activity level (light, medium, and heavy, categorised by the International Physical Activity Questionnaire) and percentage of green space coverage within 300 m buffer surrounding the participants' residential addresses. Furthermore, we applied Spearman's correlation coefficients to assess the correlations between different air pollutants. Additionally, we investigated whether the associations between air pollution and mental disorders varied by sex (female and male), age (\leq 55 years, 55 to 65 years and > 65 years), physical activity (low, medium and high-level activity), and small-area deprivation (low and high-level of Townsend deprivation index)

through interaction analyses. We also conducted stratified analyses by these factors.

To examine the possible interaction between air pollutants and PRS on the development of specific mental disorders, we created a categorical variable based on quartiles of the air pollution score and a dichotomised categorical variable for PRS (by the median). The reference group for comparison consisted of individuals with low PRS risk and exposure levels in the first quartile of the air pollution score. The Relative Excess Risk due to Interaction (RERI) was the indicator used to quantify additive interaction (Richardson and Kaufman, 2009). Zero on the RERI scale denotes the absence of additive interaction. A positive additive interaction is indicated by an RERI greater than zero. A negative additive interaction is indicated by an RERI less than zero. The significance of the interaction term in the regression model served as a proxy for the multiplicative scale of interactive effects (Yao et al., 2024). Age, sex, ethnicity, family income, educational attainment, employment status, duration of residence at the current address, Townsend deprivation index, physical activity level, percentage of green space coverage within 300 m buffer at the participants' addresses, and the leading 10 genetic principal components (PCs) were adjusted.

2.6. Sensitivity analysis

First, we restricted the diagnoses of any mental disorder to 3 years after the baseline survey to minimise the reverse causality effect on the observed associations. Second, we conducted a sensitivity analysis for participants who lived in the baseline address for at least five years to assess whether the relocation affected the main findings of air pollution and mental disorders. Third, we further adjusted for the average 24-h noise level (dB), which was modelled at the year 2009 with a version of the CNOSSOS-EU model (Kephalopoulos et al., 2014; Kephalopoulos et al., 2012). Fourth, we estimated the effects of the air pollution score on incident mental disorders with the participants whose PRSs for schizophrenia, depression and anxiety were available. Finally, to account for the influence of small-area climate factors (Thompson et al., 2023; Li et al., 2023), we controlled for the clustering effects of the 22 UK Biobank assessment centres as a sensitivity analysis. These centres are distributed across Scotland, England, and Wales, covering a majority of the cities in the UK. Participants took the baseline assessment at the assessment centre within 40 km of their residential address (Fry et al., 2017).

We used R software (version 4.3.1) for statistical analyses, with the *"survival"* package employed to analyse the Cox regression and the *"interactionR"* package used to calculate the interaction effects.

3. Results

We included 252,376 participants in the current study to assess the association between residential air pollution and incident mental disorders. Characteristics of participants with and without complete covariate information were shown in Supplementary Table 1. It indicated that there was no significant difference between the participants before and after the exclusion (Cohen, 1988; Austin, 2009). The characteristics of the analytic population were described (Table 1). The participants have lived at their current addresses for 17.21 ± 11.71 (mean \pm SD) years. The participants' average age was 56.19 ± 8.14 years, with 90.9% being White. The average percentage of green space coverage within a 300-m buffer surrounding the residential address was 35.52 ± 23.58 . The distribution of air pollutants was shown in Fig. 2. Air pollutants' concentrations were correlated, with high correlations (r > 0.8) observed between PM_{2.5} absorbance and NO₂ (r = 0.84), PM_{2.5} and NO₂ (r = 0.85), PM_{2.5} and NO_x (r = 0.88), and NO₂ and NO_x (r = 0.92).

After a median of 13.36 (IQR: 12.51–14.10) years follow-up, we observed 35,819 incidents of mental disorders, with 358 incidents of schizophrenia, 9586 incidents of depression and 10,227 incidents of anxiety. The adjusted and crude associations between air pollution and

Table 1

Characteristics of the study population at baseline (2006-2010).

	N (%)
	252,376
Age, years, Mean (SD)	56.19 (8.14)
Sex	
Female	125,671 (49.8)
Male	126,705 (50.2)
Ethnicity	
White	229,533 (90.9)
Non-white	22,843 (9.1)
Length of living in current address, Mean (SD)	17.21 (11.71)
Qualification	
University or college	93,915 (37.2)
A-levels or equivalent	30,255 (12.0)
GCSEs or equivalent	67,373 (26.7)
Other	60,833 (24.1)
Household income	
Less than £18,000	48,719 (19.3)
£18,000 to £30,999	62,563 (24.8)
£31,000 to £51,999	67,737 (26.8)
£52,000 to 100,000	57,069 (22.6)
>100,000	16,288 (6.5)
Employment status	
Employed	164,604 (65.2)
Retired	77,494 (30.7)
Unemployed	10,278 (4.1)
Townsend Deprivation Index	
Qn1 (least deprived)	50,490 (20.0)
Qn2	50,462 (20.0)
Qn3	50,475 (20.0)
Qn4	50,475 (20.0)
Qn5 (most deprived)	50,474 (20.0)
Physical activity	
Low	45,851 (18.2)
Medium	103,493 (41.0)
High	103,032 (40.8)
Green space coverage, Mean (SD)	35.52 (23.58)

SD: standard deviation; Qn: Quantile.

incident mental disorders were shown in Table 2 and Supplementary Table 2. Table 2 shows that per IQR increases in exposure to PM_{2.5} absorbance (0.32 per meter), $PM_{2.5}$ (1.28 µg/m³), NO_2 (10.08 µg/m³), and NO_x (16.78 μ g/m³) were associated with a higher risk of incident overall mental disorders ($PM_{2.5}$ absorbance: HR = 1.02, 95 % CI: 1.01–1.04; $PM_{2.5}$: HR = 1.06, 95 % CI: 1.04–1.08; NO_2 : HR = 1.06, 95 % CI: 1.04–1.08; and NO_x: HR = 1.04, 95 % CI: 1.03–1.06). Specifically, PM_{2.5} absorbance, NO₂, and NO_x were associated with a higher risk of schizophrenia (PM_{2.5} absorbance: HR = 1.18, 95 % CI: 1.05–1.33; NO₂: HR = 1.29, 95 % CI: 1.09–1.52; and NO_x : HR = 1.16, 95 % CI: 1.04-1.29). Per IQR increase in PM2.5. NO2 and NOx were correlated with an increased risk of depression ($PM_{2.5}$: HR = 1.04, 95 % CI:1.01–1.08; NO₂: HR = 1.04, 95 % CI:1.00–1.08; and NO_x: HR = 1.03, 95 % CI:1.00–1.05) and anxiety (PM_{2.5}: HR = 1.07, 95 % CI:1.03–1.10; NO_2: HR = 1.07, 95 % CI:1.03–1.11; and NO_x: HR = 1.04, 95 % CI:1.02-1.07). PM_{2.5-10} and PM₁₀ were not associated with overall or specific mental disorders.

The associations between air pollution scores and incident mental disorders were described in Table 3. In comparison to the first quartile of air pollution score, the second (HR = 1.05, 95 % CI: 1.01–1.18), third (HR = 1.13, 95 % CI: 1.09–1.18) and fourth (HR = 1.14, 95 % CI: 1.09–1.19) quartile of the score were related to a larger risk of incident mental disorders. Similar associations were observed between the air pollution score and incident anxiety. Participants with exposure in the 3rd and 4th quartile of air pollution score were associated with an increased risk of depression (3rd quartile: HR = 1.12, 95 % CI: 1.04–1.20; 4th quartile: HR = 1.10, 95 % CI: 1.02–1.19), compared to the 1st quartile. No associations were observed for the quartile groups of air pollution score and schizophrenia. In addition, an interquartile increase in the air pollution score was associated with 6 % (HR = 1.06, 95

Air Pollutants	Minimum	1st quartile	Mean	3rd quartile	Maximum	PM _{2.5} absorbance	PM _{2.5}	PM _{2.5-10}	PM ₁₀	NO ₂	NOx
PM _{2.5} absorbance	0.83	0.99	1.19	1.31	4.60	1.00					
PM _{2.5}	8.17	9.26	9.96	10.54	21.31	0.63	1.00				
PM _{2.5-10}	5.57	5.84	6.42	6.62	12.82	0.45	0.28	1.00			
PM ₁₀	11.78	15.22	16.20	16.99	31.39	0.57	0.58	0.77	1.00		
NO ₂	12.93	21.18	26.55	31.26	108.49	0.84	0.85	0.30	0.55	1.00	
NOx	19.74	33.82	43.67	50.60	265.94	0.73	0.88	0.30	0.56	0.92	1.00

Fig. 2. Air pollutant concentrations and their correlations

Footnotes: The units for $PM_{2.5}$, $PM_{2.5-10}$, PM_{10} , NO_2 , and NO_x are $\mu g/m^3$. The unit for $PM_{2.5}$ absorbance is per metre.

Table 2
Associations between individual air pollutants and the risk of incident mental disorders.

Air pollutants		Mental disorders C	ases = 35,819	Schizophrenia Case	es = 358	Depression Cases =	9568	Anxiety Cases = $10,227$		
		Hazard ratio	P value	Hazard ratio	P value	Hazard ratio	P value	Hazard ratio	P value	
PM _{2.5} Absorbance	1st quartile	Reference		Reference		Reference		Reference		
	2nd quartile	1.06 (1.02, 1.09)	0.001	1.03 (0.74, 1.43)	0.864	1.08 (1.01, 1.14)	0.017	1.10 (1.03, 1.16)	0.002	
	3rd quartile	1.10 (1.06, 1.14)	< 0.001	1.16 (0.82, 1.64)	0.393	1.08 (1.02, 1.16)	0.014	1.17 (1.10, 1.25)	< 0.001	
	4th quartile	1.08 (1.04, 1.12)	< 0.001	1.39 (0.98, 1.98)	0.064	1.05 (0.98, 1.12)	0.203	1.09 (1.02, 1.16)	0.012	
	P for trend		< 0.001		0.032		0.498		0.035	
	Per IQR increase	1.02 (1.01, 1.04)	0.002	1.18 (1.05, 1.33)	0.006	1.01 (0.98, 1.04)	0.483	1.02 (1.00, 1.05)	0.072	
PM _{2.5}	1st quartile	Reference		Reference		Reference		Reference		
	2nd quartile	1.06 (1.03, 1.10)	0.001	1.44 (0.99, 2.10)	0.057	1.07 (1.00, 1.14)	0.068	1.13 (1.06, 1.21)	< 0.001	
	3rd quartile	1.10 (1.06, 1.14)	< 0.001	1.51 (1.00, 2.28)	0.047	1.11 (1.03, 1.19)	0.008	1.15 (1.07, 1.24)	< 0.001	
	4th quartile	1.13 (1.09, 1.18)	< 0.001	1.85 (1.21, 2.84)	0.005	1.13 (1.04, 1.22)	0.003	1.21 (1.12, 1.31)	< 0.001	
	P for trend		< 0.001		0.007		0.003		< 0.001	
	Per IQR increase	1.06 (1.04, 1.08)	< 0.001	1.14 (0.98, 1.34)	0.085	1.04 (1.01, 1.08)	0.009	1.07 (1.03, 1.10)	< 0.001	
PM _{2.5-10}	1st quartile	Reference		Reference		Reference		Reference		
	2nd quartile	1.01 (0.98, 1.04)	0.344	0.84 (0.62, 1.14)	0.266	0.95 (0.89, 1.00)	0.058	1.03 (0.97, 1.08)	0.360	
	3rd quartile	1.02 (0.99, 1.05)	0.144	0.97 (0.72, 1.31)	0.844	0.96 (0.90, 1.01)	0.143	1.02 (0.97, 1.08)	0.441	
	4th quartile	1.00 (0.97, 1.03)	0.888	1.06 (0.79, 1.41)	0.711	0.96 (0.90, 1.01)	0.115	1.00 (0.95, 1.06)	0.868	
	P for trend		0.779		0.350		0.324		0.838	
	Per IQR increase	1.00 (0.99, 1.01)	0.968	1.01 (0.92, 1.10)	0.904	1.00 (0.99, 1.02)	0.588	1.00 (0.99, 1.02)	0.657	
PM10	1st quartile	Reference		Reference		Reference		Reference		
	2nd quartile	1.01 (0.98, 1.05)	0.392	1.19 (0.85, 1.67)	0.307	1.00 (0.94, 1.07)	0.987	1.07 (1.01, 1.14)	0.034	
	3rd quartile	1.05 (1.01, 1.09)	0.006	1.11 (0.78, 1.58)	0.573	1.04 (0.97, 1.11)	0.241	1.11 (1.04, 1.18)	0.002	
	4th quartile	1.00 (0.97, 1.04)	0.896	1.15 (0.82, 1.61)	0.430	0.98 (0.92, 1.05)	0.598	1.05 (0.98, 1.12)	0.141	
	P for trend		0.883		0.597		0.540		0.321	
	Per IQR increase	1.00 (0.99, 1.01)	0.613	1.02 (0.91, 1.13)	0.751	1.01 (0.99, 1.03)	0.370	1.02 (1.00, 1.04)	0.131	
NO ₂	1st quartile	Reference		Reference		Reference		Reference		
	2nd quartile	1.07 (1.04, 1.11)	< 0.001	1.07 (0.74, 1.55)	0.717	1.10 (1.03, 1.18)	0.004	1.09 (1.02, 1.16)	0.008	
	3rd quartile	1.13 (1.09, 1.18)	< 0.001	1.26 (0.85, 1.87)	0.243	1.14 (1.06, 1.23)	< 0.001	1.19 (1.11, 1.27)	< 0.001	
	4th quartile	1.16 (1.11, 1.21)	< 0.001	1.62 (1.06, 2.48)	0.027	1.11 (1.03, 1.21)	0.009	1.19 (1.10, 1.29)	< 0.001	
	P for trend		< 0.001		0.010		0.024		< 0.001	
	Per IQR increase	1.06 (1.04, 1.08)	< 0.001	1.29 (1.09, 1.52)	0.003	1.04 (1.00, 1.08)	0.030	1.07 (1.03, 1.11)	< 0.001	
NO _x	1st quartile	Reference		Reference		Reference		Reference		
	2nd quartile	1.06 (1.02, 1.09)	0.002	1.10 (0.76, 1.59)	0.604	1.06 (0.99, 1.13)	0.107	1.11 (1.04, 1.18)	0.002	
	3rd quartile	1.11 (1.07, 1.16)	< 0.001	1.18 (0.80, 1.75)	0.399	1.08 (1.00, 1.16)	0.043	1.20 (1.12, 1.28)	< 0.001	
	4th quartile	1.15 (1.11, 1.20)	< 0.001	1.63 (1.09, 2.45)	0.017	1.10 (1.02, 1.19)	0.011	1.22 (1.14, 1.32)	< 0.001	
	P for trend		< 0.001		0.005		0.015		< 0.001	
	Per IQR increase	1.04 (1.03, 1.06)	< 0.001	1.16 (1.04, 1.29)	0.006	1.03 (1.00, 1.05)	0.047	1.04 (1.02, 1.07)	0.001	

Adjusted for sex, gender, ethnicity, duration of residence at the current address, income level, education qualifications, employment status, activity group, Townsend deprivation index, and green space coverage in a 300 m buffer of residential addresses. The units for $PM_{2.5}$, $PM_{2.5-10}$, PM_{10} , NO_2 and NO_x are $\mu g/m^3$, and the unit for $PM_{2.5}$ absorbance is per meter.

% CI: 1.04–1.07), 24 % (HR = 1.24, 95 % CI: 1.08–1.43), 4 % (HR = 1.04, 95 % CI: 1.01–1.07) and 6 % (HR = 1.06, 95 % CI: 1.03–1.10) larger risk of incident mental disorder, schizophrenia, depression and anxiety, respectively. Age and deprivation levels varied the associations between air pollution scores and mental disorders, depression and

anxiety (Fig. 3).

The distributions of PRS between participants with and without schizophrenia, depression and anxiety were shown in Supplementary Fig. 2. The associations between PRS decile and the presence of schizophrenia, depression and anxiety until the end of follow-up were

Table 3

Associations between air pollution score and risk of incident mental disorders.

	Mental disorders		Schizophrenia		Depression		Anxiety	
	Hazard Ratio	P value						
1st quartile	Reference		Reference		Reference		Reference	
2nd quartile	1.05 (1.01, 1.08)	0.010	1.10 (0.77, 1.59)	0.597	1.04 (0.97, 1.11)	0.305	1.13 (1.06, 1.21)	< 0.001
3rd quartile	1.13 (1.09, 1.18)	< 0.001	1.18 (0.80, 1.74)	0.410	1.12 (1.04, 1.2)	0.003	1.25 (1.16, 1.34)	< 0.001
4th quartile	1.14 (1.09, 1.19)	< 0.001	1.42 (0.94, 2.15)	0.094	1.10 (1.02, 1.19)	0.018	1.21 (1.12, 1.31)	< 0.001
P for trend		< 0.001		0.064		0.013		< 0.001
Per IOR increase	1.06 (1.04, 1.07)	< 0.001	1.24 (1.08, 1.43)	0.003	1.04 (1.01, 1.07)	0.022	1.06 (1.03, 1.10)	< 0.001

Adjusted for sex, gender, ethnicity, duration of residence at the current address, income level, education qualifications, employment status, activity group, Townsend deprivation index, and green space coverage in a 300 m buffer of residential addresses.

Mental disorders

group	Ν	estimation		p-value	p-value for interaction term
All subjects	252376	1.06 (1.04, 1.07)		<0.001	
Sex					0.453
Female	125671	1.05 (1.03, 1.08)		<0.001	
Male	126705	1.06 (1.04, 1.08)		<0.001	
Age					0.001
<= 55	110400	1.03 (1.01, 1.06)		0.018	
55 - 65	107306	1.07 (1.04, 1.10)		<0.001	
> 65	34670	1.08 (1.04, 1.12)		<0.001	
Physical activity					0.270
low-level activity	45851	1.06 (1.02, 1.10)		0.002	
medium-level activity	103493	1.07 (1.04, 1.10)		<0.001	
high-level activity	103032	1.05 (1.02, 1.07)		0.001	
Townsend deprivation index					0.002
low-level deprivation	126208	1.12 (1.08, 1.15)		<0.001	
High-level deprivation	126168	1.08 (1.06, 1.10)	-	<0.001	
		~	0.95 1 1.1 1.2		

Protective Adverse

group	Ν	estimation		p-value	p-value for interaction term
All subjects	252376	1.24 (1.08, 1.43)		0.003	
Sex					0.156
Female	125671	1.38 (1.14, 1.67)		0.001	
Male	126705	1.11 (0.89, 1.37)		0.346	
Age					0.961
<= 55	110400	1.26 (1.00, 1.59)		0.053	
55 - 65	107306	1.23 (0.98, 1.54)		0.070	
> 65	34670	1.26 (0.92, 1.73)	-	0.154	
Physical activity					0.352
low-level activity	45851	1.25 (0.90, 1.72)		0.185	
medium-level activity	103493	1.17 (0.92, 1.48)		0.195	
high-level activity	103032	1.30 (1.05, 1.62)		0.018	
Townsend deprivation index					0.314
low-level deprivation	126208	1.14 (0.83, 1.58)		0.418	
High-level deprivation	126168	1.32 (1.13, 1.53)			
			0.91 1.2 1.4 1.	6 1.8	

Protective Adverse

Depression

					Anxiety					
N	estimation		p-value	p-value for interaction term	group	N	estimation		p-value	p-value for interaction term
252376	1.04 (1.01, 1.07)		0.022		All subjects	252376	1.06 (1.03, 1.10)		< 0.001	
		1		0.134	Sex					0.614
125671	1.02 (0.98, 1.07)		0.289		Female	125671	1.07 (1.03, 1.11)		0.001	
126705	1.06 (1.01, 1.11)		0.023		Male	126705	1.05 (1.00, 1.11)		0.043	
				0.009	Age			1		0.001
110400	0.99 (0.95, 1.04)		0.776		<= 55	110400	1.00 (0.95, 1.05)		0.940	
107306	1.06 (1.01, 1.11)		0.024		55 - 65	107306	1.10 (1.05, 1.15)		<0.001	
34670	1.12 (1.04, 1.21)		-0.005		> 65	34670	1.15 (1.07, 1.24)		-<0.001	
		1		0.080	Physical activity			i i		0.732
45851	1.03 (0.96, 1.10)		0.411		low-level activity	45851	1.03 (0.96, 1.10)		0.460	
103493	1.03 (0.98, 1.08)		0.327		medium-level activity	103493	1.10 (1.04, 1.15)		<0.001	
103032	1.06 (1.01, 1.11)		0.030		high-level activity	103032	1.05 (1.00, 1.10)		0.058	
x				0.031	Townsend deprivation inde	ex				0.003
126208	1.11 (1.05, 1.18)	· · · · · ·	<0.001		low-level deprivation	126208	1.16 (1.09, 1.22)		<0.001	
126168	1.04 (1.00, 1.08)		0.033		High-level deprivation	126168	1.05 (1.01, 1.09)		0.015	
		\longrightarrow	1.2				←	\longrightarrow		
	252376 125671 126705 110400 107306 34670 45851 103493 103032 x 126208	252376 1.04 (1.01, 1.07) 125671 1.02 (0.98, 1.07) 126705 1.06 (1.01, 1.11) 110400 0.99 (0.95, 1.04) 107306 1.06 (1.01, 1.11) 34670 1.12 (1.04, 1.21) 45851 1.03 (0.96, 1.10) 103493 1.03 (0.98, 1.08) 103032 1.06 (1.01, 1.11) x 126208 1.11 (1.05, 1.18) 126208 1.04 (1.00, 1.08) _	252376 1.04 (1.01, 1.07) 125671 1.02 (0.98, 1.07) 126705 1.06 (1.01, 1.11) 110400 0.99 (0.95, 1.04) 107306 1.06 (1.01, 1.11) 34670 1.12 (1.04, 1.21) 45851 1.03 (0.96, 1.10) 103493 1.03 (0.98, 1.08) 103302 1.06 (1.01, 1.11) * * 126208 1.11 (1.05, 1.18) 1262108 1.04 (1.00, 1.08)	252376 1.04 (1.01, 1.07) - 0.022 125671 1.02 (0.98, 1.07) - 0.289 126705 1.06 (1.01, 1.11) - 0.023 110400 0.99 (0.95, 1.04) - 0.776 107306 1.06 (1.01, 1.11) - 0.024 34670 1.12 (1.04, 1.21) - 0.025 45851 1.03 (0.96, 1.10) - 0.411 103032 1.08 (1.01, 1.11) - 0.327 103032 1.06 (1.01, 1.11) - 0.030 x - - - 0.033 126208 1.11 (1.05, 1.18) - - 0.033 0.95 1.1 1.2 - - -	N estimation p-value interaction term 252376 1.04 (1.01, 1.07) - 0.022 0.134 126671 1.02 (0.98, 1.07) - 0.289 0.134 126675 1.06 (1.01, 1.11) 0.023 0.009 110400 0.99 (0.95, 1.04) - 0.776 107306 1.06 (1.01, 1.11) - 0.024 34670 1.12 (1.04, 1.21) - 0.005 45851 1.03 (0.96, 1.10) - 0.411 103032 1.06 (1.01, 1.11) - 0.030 x 0.031 - 0.031 126208 1.11 (1.05, 1.18) - - 0.95 1.11 1.2 -	N estimation p-value interaction term group 252376 1.04 (1.01, 1.07) 0.022 0.134 All subjects Sex Female Male M	N estimation p-value interaction term 252376 1.04 (1.01, 1.07) 0.022 All subjects 252376 126671 1.02 (0.98, 1.07) 0.289 Female 125671 126670 1.06 (1.01, 1.11) 0.023 6.009 General 6.65 110400 0.99 (0.95, 1.04) 0.776 \sim 55 110400 55 - 65 107306 34670 1.12 (1.04, 1.21) 0.005 0.080 5 - 65 107306 45851 1.03 (0.96, 1.10) 0.411 0.030 5 - 65 34670 103493 1.03 (0.98, 1.08) 0.327 medium-level activity 103323 104302 1.11 (1.05, 1.18) 0.033 103032 Townsend deprivation index 126208 1.104 (1.00, 1.08) 0.033 0.033 122 124	N estimation p-value interaction term 252376 1.04 (1.01, 1.07) 0.022 126671 1.02 (0.98, 1.07) 0.229 126675 1.06 (1.01, 1.11) 0.023 126705 1.06 (1.01, 1.11) 0.024 10400 0.99 (0.95, 1.04) 0.076 107306 1.06 (1.01, 1.11) 0.024 34670 1.12 (1.04, 1.21) 0.005 45851 1.03 (0.96, 1.10) 0.411 103493 1.03 (0.96, 1.10) 0.411 103493 1.03 (0.98, 1.08) 0.327 103322 1.06 (1.01, 1.11) 0.031 126208 1.11 (1.05, 1.18) 0.031 126208 1.11 (1.05, 1.18) 0.033 126208 1.10 (1.00, 1.08) 0.033 0.955 1.11 1.2	N estimation p-value interaction term 252376 1.04 (1.01, 1.07) $-$ 0.022 126671 1.02 (0.98, 1.07) $-$ 0.289 126675 1.06 (1.01, 1.11) $-$ 0.024 107006 1.06 (1.01, 1.11) $-$ 0.009 0.009 0.99 (0.95, 1.04) $-$ 0.76 107306 1.06 (1.01, 1.11) $-$ 0.024 34670 1.12 (1.04, 1.21) $-$ 0.060 45851 1.03 (0.96, 1.10) $ -$ 45851 1.03 (0.96, 1.10) $ -$ 103493 1.03 (0.98, 1.10) $ -$ 103392 1.06 (1.01, 1.11) $ 0.031$ 126208 1.11 (1.05, 1.18) $ -$ 126208 1.11 (1.05, 1.18) $ 0.951$ 1.11 1.2 $-$	N estimation p-value interaction term group N estimation p-value 252376 1.04 (1.01, 1.07) - 0.022 0.134 $ -$ </td

Anviety

Fig. 3. Stratification analyses for the association of an interquartile range increase in air pollution score and risk of incident mental disorders.

demonstrated in Supplementary Fig. 3, with the first decile as the reference group and adjustment of age, sex, and the primary ten genetic PCs. People in the highest PRS decile had 414 % higher risk of schizophrenia (odds ratio = 5.14, 95 % CI: 3.88, 6.83), 79 % higher risk of depression (odds ratio = 1.79, 95 % CI: 1.70, 1.89), and 20 % higher risk of anxiety (odds ratio = 1.20, 95 % CI: 1.13, 1.28) than those in the lowest decile, respectively. Participants with a high genetic risk were associated with a higher risk of incident mental disorders regardless of air pollution exposure (Fig. 4). Neither additive nor multiplicative interaction effects between air pollutant exposure and PRS was found on incident schizophrenia, depression, or anxiety.

A series of sensitivity analyses showed consistent results with our main analyses. The associations between air pollution score and incident mental disorders remained similar after excluding participants who were diagnosed with mental disorders within three years of recruitment (Supplementary Table 3) or excluding those who have stayed at their current address for <5 years (Supplementary Table 4). The associations between air pollution score and incident mental disorders remained

e increase in air pollution score and risk of incident mental disorders. comparable to the main analysis when we additionally adjusted for 24-h average noise (Supplementary Table 5), or when we limited to participants whose PRS was available (Supplementary Table 6), or when we

additionally adjusted for small-area climate factors (Supplementary

4. Discussion

Table 7).

Based on a large-scale and long-term prospective cohort, we investigated the relationship between exposure to air pollutants and the incident mental disorders, including schizophrenia, depression, and anxiety. Our findings revealed that PM_{2.5}, NO₂, and NOx were risk factors for incident mental disorders, depression, and anxiety. By constructing an innovative composite score of air pollutants, we demonstrated a robust positive relationship between multiple air pollutants and the risk of mental disorders. Additionally, we explored the interaction between air pollution and genetic susceptibility. To the best of our knowledge, this is the first study to provide evidence on whether

Air pollutantion score	Genetic risk	Hazard Ratio		Additive effect RERI	Multiplicative effect P value
Schizophrenia	Ū.				
Q1	Low	reference	+	-	-
	High	2.46 (1.26, 4.83)			
Q2	Low	1.37 (0.63, 3.00)	· · · · · · · · · · · · · · · · · · ·	-0.27 (-1.91, 1.37)	0.551
	High	2.56 (1.26, 5.19)			
Q3	Low	1.26 (0.55, 2.90)	· · · · · ·	-0.14 (-1.78, 1.49)	0.693
	High	2.58 (1.23, 5.40)			
Q4	Low	2.12 (0.95, 4.72)	+	-0.98 (-2.91, 0.96)	0.117
	High	2.61 (1.20, 5.64)	•		
Depression			0.5 2.5 4.5		
Q1	Low	reference	•		
X -	High	1.25 (1.11, 1.40)			
Q2	Low	1.00 (0.88, 1.13)		0.08 (-0.10, 0.25)	0.451
~ -	High	1.32 (1.17, 1.49)		0.000 (0.000, 0.000)	
Q3	Low	1.16 (1.02, 1.32)	· · · · · · · · · · · · · · · · · · ·	-0.03 (-0.21, 0.16)	0.548
	High	1.38 (1.22, 1.56)			
Q4	Low	1.12 (0.97, 1.28)	÷	0.04 (-0.14, 0.23)	0.914
	High	1.40 (1.23, 1.60)			
Anxiety			0.9 1.1 1.3 1.5		
Q1	Low	reference			
QI			•		
02	High	1.04 (0.94, 1.16)		0.01 (0.15 0.18)	0.057
Q2	Low	1.17 (1.05, 1.32)		0.01 (-0.15, 0.18)	0.957
01	High	1.23 (1.10, 1.38)		0.01 (0.18, 0.16)	0.925
Q3	Low	1.25 (1.11, 1.41)	1	-0.01 (-0.18, 0.16)	0.825
04	High	1.29 (1.14, 1.45)		0.00 (0.17 0.10)	0.029
Q4	Low	1.27 (1.12, 1.45)		0.00 (-0.17, 0.18)	0.938
	High	1.32 (1.16, 1.50)			
			0.9 1.1 1.3 1.5		

Fig. 4. Interaction of air pollution and genetic risk on specific mental disorders (N = 139,802).

Footnotes: Adjusted for sex, gender, ethnicity, length of living at the current address, income level, education qualifications, employment status, activity group, Townsend deprivation index, green space coverage at a 300 m buffer of residential addresses, polygenetic risk scores (PRS) of specific mental disorders (including schizophrenia, depressions, and anxiety, respectively) and the first 10 genetic principal components.

Genetic risk was categorised into two levels by median of PRS. Air pollution score was categorised into four levels by quartile. We used the first quartile of the air pollution score and the low PRS as the reference group to calculate additive and multiplicative effects for the other seven groups.

The *p*-values of the interaction of air pollution score (as a linear term) and genetic risk (as a linear term) on specific mental disorders were 0.946, 0.439 and 0.742 for schizophrenia, depression and anxiety, respectively.

genetic susceptibility modifies the relationship between concurrent exposure to various air pollutants and incident mental disorders. We observed that although no additive or multiplicative interaction of PRS and air pollution scores on incident schizophrenia, depression or anxiety was found, the associations between air pollution and schizophrenia, depression and anxiety consistently showed a trend of enhancement in the group with a high genetic predisposition.

Studies investigating the impact of a single air pollutant on mental well-being have been widely reported. However, because of inconsistencies in outcome identification and study design, there's no consensus on the link between air pollution and specific mental illnesses such as depression, anxiety or schizophrenia (Braithwaite et al., 2019; Fan et al., 2020; Antonsen et al., 2020; Zhang et al., 2024). Our study demonstrated that long-term exposure to PM_{2.5} was associated with a higher risk of mental disorders, consistent with previous studies (Braithwaite et al., 2019; Zhang et al., 2024; Borroni et al., 2022). NO₂,

NOx and PM_{2.5} absorbance, a surrogate indicator of elemental carbon, are principally released from road transport in the UK (Lyons et al., 2024; Zeng et al., 2019; Eeftens et al., 2012b). Our findings support the findings seen in other studies that exposure to a certain higher level of traffic-related air pollutants could result in a higher risk of mental disorders (Nobile et al., 2023; Yang et al., 2023; Zare Sakhvidi et al., 2022). The association was evident after further adjustment of traffic-related noise (Supplementary Table 5). Different from previous studies using the same cohort, our studies did not observe adverse associations between PM₁₀ and incident depression, anxiety or schizophrenia (Zhang et al., 2024; Gao et al., 2023). One possible explanation for this discrepancy is that we only used the 2010 air pollutant measurements, not the average concentrations for 2007 and 2010. As different regression models for air pollutant estimation were used for each of these two years (air pollution estimates for the year 2007 were derived using the EU-wide air pollution maps instead of the LUR), UK Biobank did not

consider it appropriate to average concentrations for these two years (https://biobank.ctsu.ox.ac.uk/ukb/label.cgi?id=114). Furthermore, the aforementioned study employed the depression screening questionnaire (9-item Patient Health Questionnaire) to determine the outcome of depression rather than using the ICD-10 for clinical diagnosis. This might lead to an overestimation of the associations (Thombs et al., 2018). Nevertheless, our findings were consistent with those of previous research conducted in Europe that indicated long-term exposure to PM₁₀ had no significant associations with depression, anxiety or schizophrenia (Zeng et al., 2019; Antonsen et al., 2020; Petrowski et al., 2021). We also observed similar findings to this study in previous research on the health effects of $PM_{2.5-10}$, with no significant adverse effects on depression. This suggests that the mental health implications of PM_{2.5-10} may be less threatening than those for PM_{2.5}, which were similar to those findings observed in other health outcomes (Adar et al., 2014; Zhang et al., 2023). In agreement with earlier investigations using air pollution scores, our study, which had a longer follow-up period, covered more types of air pollutants and defined study outcomes based on clinician-diagnosed episodes of mental disorders, supporting the hypothesis that concurrent air pollutants exposure could lead to a higher risk of mental disorders (Zhang et al., 2024; Yang et al., 2023; Gao et al., 2023).

Previous research has explored the underlying mechanisms that could elucidate the associations between exposure to air pollutants and the occurrence of mental disorders. Neuroinflammation and oxidative stress are well-recognised as common mechanisms of air pollution that contribute to mental disorders (Antonsen et al., 2020; Block and Calderón-Garcidueñas, 2009; Power et al., 2015; van den Bosch and Meyer-Lindenberg, 2019; Sawa and Sedlak, 2016; Vogelzangs et al., 2013). The particulate matter, especially small particles, would reach the brain and induce inflammation and oxidative stress in the central nervous system, and nitrogen oxides would initiate an oxidative signalling cascade that further induces neuroinflammation (Lundberg and Weitzberg, 2022; Adar et al., 2014). However, distinct psychiatric diseases may involve different biological pathways and exhibit varying susceptibility to air pollutants. For instance, inhaling air pollutants can activate the hypothalamic-pituitary-adrenal (HPA) axis, increasing stress hormone levels and leading to dopaminergic neurotoxicity, which raises the risk of depression and anxiety (Thomson, 2019; Dantzer et al., 2008; Hasler, 2010; Braithwaite et al., 2019; Thomson et al., 2013; Li et al., 2017; Lim et al., 2012). As a neurodevelopmental disorder (Antonsen et al., 2020), schizophrenia may exhibit greater susceptibility to air pollution compared to depression or anxiety. Previous studies have indicated that air pollution can induce neurotoxicity, disrupt neurodevelopment, and lead to neurological dysfunction, thereby increasing the risk of schizophrenia. Specifically, inflammation caused by air pollution can interfere with glutamate-mediated neurotransmission and induce neurotoxic effects, further contributing to the pathogenesis of schizophrenia (Marsman et al., 2013; Moghaddam and Javitt, 2012). Long-term exposure to air pollutants during early life can disrupt normal brain development, thereby elevating the likelihood of developing schizophrenia in later stages (Antonsen et al., 2020). Additionally, evidence suggests that prolonged exposure to air pollution is associated with cognitive decline and other neurological impairments, which may exacerbate the severity of schizophrenia (McCutcheon et al., 2023; Schikowski and Altuğ, 2020). Our findings showed that air pollution's adverse effects were more health-threatening among older adults. Older adults are more susceptible to inflammation triggered by substances such as air pollutants due to increased vulnerability of the immune system and nervous system (Qiu et al., 2023; Shumake et al., 2013). Associations between air pollution and mental disorders (e.g. depression and anxiety) possibly varied by area deprivation (Brunt et al., 2016).

Studying the interaction between genetic susceptibility and the environment may help explain disease development (Peters et al., 2021; Jin et al., 2020). Gene-environment interaction studies using a longitudinal study design are not sufficient. Our study provides research

evidence on the interaction of genetic predisposition and multiple air pollutants on incident schizophrenia, anxiety and depression. The results implied that mental disorders might be affected by genes and air pollutants in separate pathways. In one previous research investigating the interaction between genetic predisposition and joint effects of multiple air pollutants on mental disorders (depression and anxiety), no evidence was found for either additive or multiplicative geneenvironment interactions (Gao et al., 2023). Another geneenvironment interaction study observed an additive joint interaction between individual air pollutants and genetic factors on the risk of depression (Fu et al., 2022). In another interaction effects study investigating the genetic effects of schizophrenia and four specific air pollutants (PM2.5, PM10, NO2, NOx and joint effect) on late-onset incident schizophrenia, additive effects were observed among the highest level of genetic risk (categorised by tertile) and the high level of air pollutant scores (categorised by median) (Zhang et al., 2024). Our study found no multiplicative nor additive gene-environment interactions on mental disorders, including schizophrenia, depression and anxiety. In this study, we considered the combined effects of various air pollutants instead of a single pollutant and covered wider composites of air pollution. In addition, we assessed the PRS based on summary statistics from meta-analyses excluding UK Biobank participants. This minimises the overfitting concern in PRS calculation introduced by sample overlap between base and target data. We also considered a broader range of SNPs from the White ethnic samples, which better reflects the genetic basis of the mental disorders (Wray et al., 2013; Collister et al., 2022). Nevertheless, despite the absence of interaction between PRS and air pollution scores on incident mental disorders in our study, the positive correlations between air pollution and these specific mental disorders are on the rise in people with high genetic susceptibility.

This study used a large-scale cohort with a median of 13.36 years follow-up that enabled sufficient statistical power for observations of incident mental disorders. To date, limited research has examined the combined effects of particulate matter and gaseous air pollutants on the incidence of mental disorders. Given that ambient air pollution is a combination of gaseous pollutants and particulate matter, examining their combined impacts on mental illnesses is essential to illuminating the comprehensive influence of air pollution on mental health in real-world situations. We added to the previous literature by assessing the combined impacts of exposures to $PM_{2.5}$ absorbance, $PM_{2.5,}$, $PM_{2.5-10}$, PM_{10} , NO_2 and NO_x with a novel air pollution score, and found a positive correlation with the onset of mental disorders. We also carefully addressed methodological limitations in PRS calculation in previous studies and yielded the most predictive information on an individual's genetic risk for mental disorders.

However, several limitations exist in this study. Firstly, participants in the UK Biobank were recruited voluntarily. Selection bias might exist and have an impact on the representativeness and generalisability (Fry et al., 2017). Secondly, we used annual average air pollution exposure levels at the baseline and did not take into account dynamic changes in air pollution because dynamic monitoring data were limited. However, according to the follow-up surveys in the UK Biobank, participants who relocated was at a low rate of about 2 %. Since 2010, the annual emissions of particulate matter (PM10 and PM2.5) in the UK have been relatively stable, while the emissions of nitrogen oxides have shown a downward trend (Department for Environment, Food and Rural Affairs, The UK Government, 2024). Accordingly, the baseline airborne particulate matter levels could be a suitable surrogate representing long-term exposure, and we may have underestimated the association between nitrogen oxides and mental disorders. Thirdly, the average age of the participants is 56.19 years, so conclusions drawn from our study would not be generalised to younger adults. Future studies on air pollution and mental health among younger adults are suggested. Fourthly, the possibility of residual confounding, such as the climate factors (Thompson et al., 2023; Li et al., 2023), cannot be ruled out, though we have taken into account a range of important confounding variables, e.g.,

demographic features, individual and area-level socioeconomic level, duration of residence at the current address, physical activity, green space coverage, noise, and the clustering effects of assessment centres. Fifthly, the incident schizophrenia in this study should be interpreted with caution, which might be more related to late-onset schizophrenia (Shen et al., 2024). Finally, the majority of participants in this study were of European descent, and PRS was calculated based on GWAS in European descent. Therefore, caution should be exercised to generalise the findings of gene-environment interactions to different populations.

5. Conclusions

This study offered compelling evidence that long-term exposure to a variety of air pollutants was correlated to an increased risk of incident mental disorders, including schizophrenia, depression, and anxiety. Notably, we assessed the combined effects of air pollutants by constructing an innovative air pollution index, revealing the potential doseresponse to air pollution exposure and incident mental disorders. Regardless of the genetic risk of schizophrenia, depression and anxiety in this population, air pollution increased the likelihood of all and specific mental disorders. These results highlighted the importance of implementing strict air pollution control standards to reduce the disease burden of mental disorders, even in areas with milder levels of air pollution.

CRediT authorship contribution statement

Xiaoxin I. Yao: Validation, Supervision, Resources, Project administration, Methodology, Investigation, Funding acquisition, Formal analysis, Data curation, Conceptualization, Writing – original draft. Shengzhi Sun: Methodology, Writing – review & editing. Qian Yang: Methodology, Writing – review & editing. Xinning Tong: Supervision, Software, Methodology, Investigation, Formal analysis, Writing – original draft. Chen Shen: Validation, Supervision, Software, Resources, Investigation, Data curation, Writing – review & editing.

Ethics approval

All participants in this study provided written informed consent. UK Biobank was approved by the North West Multi-centre Research Ethics Committee (MREC) (Reference 16/NW/0274).

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Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Appendix A. Supplementary data

Supplementary data to this article can be found online at https://doi.org/10.1016/j.scitotenv.2025.179137.

Data availability

UK Biobank data were used under license and thus not publicly available. Access to the UK Biobank data can be requested through a standard protocol (https://www.ukbiobank.ac.uk/enable-your-res earch/apply-for-access).

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